

Radionuclides from Fukushima accident in Thessaloniki, Greece (40°N) and Milano, Italy (45°)

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ABSTRACT

¹³¹I, ¹³⁷Cs and ¹³⁴Cs were observed in environmental samples in Milano (40°N), Italy and Thessaloniki (45°N), Greece, soon after the nuclear accident in Fukushima, Japan. The radionuclide concentrations were determined and studied as a function of time. In Thessaloniki the ¹³¹I in air was observed for the first time on March 24, 2011. In Milano, the first evidence of Fukushima fallout has been confirmed with ¹³¹I and ¹³⁷Cs measured in wet precipitation collected two days later. The maximum ¹³¹I activity concentration in air of 467±25 µBq m⁻³, observed in Milano on April 3-4, 2011, was almost similar to the highest value of 497±53 µBq m⁻³ observed in Thessaloniki. The ¹³⁴Cs/¹³⁷Cs activity ratio values in air were around 1 in both regions. Soil, grass and milk samples were contaminated with ¹³¹I and ¹³⁷Cs at a low level. Finally, a dose assessment for these two areas showed clearly that the detected activities in all environmental samples were far below levels of concern.

Keywords: nuclear accident; Fukushima accident; radioactive nuclides; radioactive fallout

1. Introduction

As a consequence of the Fukushima accident, resulting from the big earthquake and Tsunami of March 2011, significant amounts of radionuclides were released to the terrestrial and marine environments [1]. Fukushima accident showed once again how fast airborne radionuclides can be transported over great distances: emissions of radionuclides were high enough to be detected all over the northern hemisphere, indicating not only local and regional but also global impact of the accident [2-13].

The most important radionuclides detected in Europe due to Fukushima accident were ¹³⁷Cs (T_{1/2}=30.6 a), ¹³⁴Cs (T_{1/2}= 2.06 a) and ¹³¹I (T_{1/2}= 8.02 d). Under normal conditions, only ¹³⁷Cs remains in the atmosphere for long time, due to its relative long half life. In Europe, ¹³⁷Cs from the Chernobyl accident was still measurable before the Fukushima accident at trace levels viz. below 1 µBq m⁻³ [14, 15], while ¹³⁴Cs due to Chernobyl accident had not been measured in the atmosphere since the middle of the 1990s.

Immediately after the Fukushima accident, monitoring activities were promptly organized worldwide. The data discussed in the present work includes the observations of Fukushima related radionuclides in Milano (45.49° N, 9.20° E), Italy by LASA (Laboratorio Acceleratori Supercoduttività Applicata) Laboratory of the INFN (Istituto Nazionale di Fisica Nucleare) of Milano and of UNIMI (Università degli Studi di Milano) and in Thessaloniki (40.50° N, 22.90° E), Greece by Nuclear Physics Laboratory of Aristotle University of Thessaloniki, giving the extent of contamination in Northern Italy and in Northern Greece due to Fukushima fallout.

The activities of ¹³¹I, ¹³⁴Cs and ¹³⁷Cs carried out in air, precipitation, soil, grass and milk at the two sites are presented here. Moreover, an evaluation of the effective dose for population was done based on the values obtained.

2. Material and Methods

Analysis for ¹³¹I, ¹³⁷Cs and ¹³⁴Cs in airborne aerosols were carried out in daily samples for both sites during March-April, 2011. Air sampling was carried out with high volume air samplers and glass fiber filters. The aerosol samples were collected for 23 hours, with one hour interval between the

collections, time interval necessary for avoiding overheating and possible destruction of the motor pump. The flow rate was about 60 cfm resulting in a total daily air volume throughput of average 2200 m³. The collection substrate was glass fiber filters type TFAGF 810 with dimension “8 x 10” with collection efficiency of 99.98% for particles as small as 0.3 µm in diameter [16].

The wet deposition was determined by exposing collection funnels to the atmosphere during the rainfall events. At the end of the rainfall, the contents of the funnels were collected and acidified with 1 mL orthophosphoric acid 85% per litre of rainwater, in order to prevent any loss by absorption to the vessel [17].

Surface soil samples were collected from a depth of 0-7 cm (upper 7 mm) and from an area of 1 m². Grass samples were collected from a surface area of 1 m² and at the same time as soil sampling.

All samples were counted using HPGe detectors at about 50% relative efficiency in case of Milano samples and with 42% efficiency in case of Thessaloniki [18, 19]. All detectors were calibrated with reference sources and standard geometries with overall uncertainties no more than 3%. In order to avoid corrections for the different geometry configuration, all samples were measured in the same geometrical assembly as the calibrated sources. In particular 1 L Marinelli geometry was used for rainwater, milk, grass and soil samples; the air filters were folded and compressed by means a hydraulic press up to 3t to obtain a cylinder of 4.7 cm in diameter and 3mm in height.

The two Laboratories performed intercomparison tests in order to assure: a) the compatibility of the sampling procedures by checking and comparing the calibration and the air samplers; b) the quality of low-background gamma spectrometry measurements, which is guaranteed by intercomparison measurements organized by international organizations like IAEA and EU. Gamma ray energies used for the analysis were 364.489 keV for the determination of ¹³¹I, 661.660 keV for ¹³⁷Cs and both 604.721 and 795.864 keV for ¹³⁴Cs [20]. In the case of ¹³⁴Cs, corrections due to coincidence effect were applied [21].

All activities given in the text and figures are decay-corrected at the mean time of the sampling interval. The overall uncertainties in the determined activities are attributed mainly to the statistical counting and sampling uncertainties, while the contribution of calibration uncertainty is less than 3%.

3. Results and Discussion

3.1. ¹³¹I, ¹³⁷Cs and ¹³⁴Cs activity concentration in surface air

In Milano region increased atmospheric radioactivity concentration for ¹³¹I (322±36 µBq m⁻³) was observed on an air filter taken on 30 March 2011, while the maximum activity concentration of 467±25 µBq m⁻³ was recorded on April 3-4. The first evidence of Fukushima fallout was confirmed with ¹³¹I and ¹³⁷Cs measured in precipitation at two sampling sites at Milano on 28 March with the concentrations of ¹³¹I and ¹³⁷Cs in the rainwater equal to 0.89±0.11 Bq L⁻¹ and 0.12±0.09 Bq L⁻¹, respectively.

The ¹³¹I was detected initially (143±21 µBq m⁻³) in Thessaloniki on 24 March 2011. The maximum ¹³¹I activity concentration of 497±53 µBq m⁻³ was detected on 4 April while a second maximum of 425±48 µBq m⁻³ on 10 April.

The time evolution of ¹³¹I activity concentration in the airborne particulate from the two regions is presented in Fig.1. It is clear that in both regions similar maximum values have been observed during 3-4 April despite the considerable distance and the possible differences in climatic parameters that may influence the concentrations of radionuclides in air, but there is a tendency of a shift in the peak days.

In Figs 2 and 3 ¹³⁷Cs and ¹³⁴Cs time evolution activity concentrations respectively in the airborne particulate for the two regions are reported. The highest values of ¹³⁷Cs and ¹³⁴Cs in air over Milano, 63±29 µBq m⁻³ and 61±22 µBq m⁻³ respectively, were observed on April 10 and over Thessaloniki, 145±18 µBq m⁻³ and 126±11 µBq m⁻³ respectively, on April 4. In most of the cases the observed maxima values for ¹³⁷Cs and ¹³⁴Cs do not follow the maxima of ¹³¹I, with the only exception is in the region of Thessaloniki on 4 April where the highest values of ¹³¹I, ¹³⁷Cs and ¹³⁴Cs were

observed. In both regions of investigation the $^{134}\text{Cs}/^{137}\text{Cs}$ ratio values remained almost stable at about 1 throughout the period of measurements. This value is in agreement with the measurements of High Energy Accelerator Research Organization at Tsukuba, Japan [22] and, in accordance with Kirchner et al. [23], it corresponds to the isotopic composition observed in Europe related to emissions of units 1 and 3 of the destroyed Fukushima reactor. As well known this ratio value is related to the burn-up history of the damaged nuclear fuel of the destroyed nuclear reactor: for comparison after the Chernobyl accident the ratio was about 0.5 [24-26].

The presence of more than one peak of ^{131}I and ^{137}Cs (Figures 1,2) demonstrates that the radioactive plume was continuously transferred from Fukushima, Japan to Europe up to end of April and/or first days of May. These results are in agreement with values for ^{131}I and ^{137}Cs reported from other European countries [4, 9 – 11, 13, 27 – 31].

The ratio of $^{131}\text{I}/^{137}\text{Cs}$ in air in Milano was 11.7 on April 3-4, reached the higher value of 16.8 on April 7-8 and then declined to 2.6 on April 28. The ratio of $^{131}\text{I}/^{137}\text{Cs}$ in air in Thessaloniki was first 18.4 on March 26 and then decreased to 3.4 on April 4. These data confirm that the radioactive cloud was first richer in ^{131}I , as iodine is a more volatile element than cesium.

After first days of May in both regions there was no more detectable presence of these artificial radionuclides. These data are consistent with the ones observed over the European continent [9] and show a rather large degree of homogeneity of the plume, following redistribution process in the troposphere.

3.2. ^{131}I , ^{137}Cs and ^{134}Cs activity concentration in rainwater samples

Wet deposition by rain is known to be the most effective transfer path for airborne radioisotopes to ground and water bodies. Wet precipitation affects the levels of deposition in the land, including grass consumed by animals that produce milk, and could affect the levels of the radioactivity in the tap water.

Convective storms on 29 March and 14 April, 2011 in Thessaloniki and 28 March and 12-15 April, 2011 in Milano were responsible for carrying the bulk of the radioactive material from the atmosphere to the ground.

The ^{131}I activity concentration in samples collected in Thessaloniki is ranging between $0.69 \pm 0.18 \text{ Bq L}^{-1}$ to $0.096 \pm 0.026 \text{ Bq L}^{-1}$ while in Milano was $0.89 \pm 0.11 \text{ Bq L}^{-1}$ and $0.72 \pm 0.13 \text{ Bq L}^{-1}$ in the samples collected in two different regions near the city of Milano on 28 March and $0.29 \pm 0.09 \text{ Bq L}^{-1}$ after the precipitation of 15 April. The observed ^{131}I values in rainwater in all the sites were below 1 Bq L^{-1} , two orders of magnitude less than the limit of 170 Bq L^{-1} [32], fixed by the Food and Drug Administration (FDA) as intervention level for ^{131}I in drinking water and infant milk. So, all tap water in Northern Greece and Italy were completely safe in terms of its ^{131}I content, taking also into account that rain water undergoes to many dilutions before becoming tap water.

3.3. ^{131}I , ^{137}Cs and ^{134}Cs activity concentration in soil, grass and milk samples

^{131}I was not detected in soil samples in Thessaloniki, while in Milano the maximum ^{131}I activity concentration observed was as high as $0.20 \pm 0.13 \text{ Bq kg}^{-1}$. No ^{134}Cs was observed from the region of Thessaloniki, while in Milano ^{134}Cs activity concentration was below 1 Bq kg^{-1} . ^{137}Cs activity concentration as high as 46 Bq kg^{-1} was observed in Thessaloniki and as high as 25 Bq kg^{-1} in Milano.

Grass was contaminated with ^{131}I and ^{137}Cs at a low level. The observed values of ^{131}I and ^{137}Cs in grass samples were in Milano between $0.037 \pm 0.013 - 0.135 \pm 0.049 \text{ Bq kg}^{-1}$ and $0.047 \pm 0.030 - 0.089 \pm 0.032 \text{ Bq kg}^{-1}$ respectively, while in Thessaloniki between $0.84 \pm 0.13 - 1.35 \pm 0.17 \text{ Bq kg}^{-1}$ and $0.33 \pm 0.11 - 0.38 \pm 0.05 \text{ Bq kg}^{-1}$ respectively. The wide range of the observed values in the two sites was probably due to the different masses collected per area as a consequence of the different growth rate at the different sampling days, the surface of the grass is not homogeneous, the leaf blades

do not present a wide surface area, while the deposited particles on the grass cannot remain attached to the leaves due to the wind, that was present in that days.

Iodine-131 was detected and measured in fresh goat and cow milk from rural areas in both sites. The highest value of ^{131}I activity concentration, 1.2 Bq L^{-1} , was observed in Thessaloniki on 30 March. The reported ^{131}I values in milk samples in Milano were below 1 Bq L^{-1} . After the end of May there was no more detectable presence of ^{131}I in milk samples.

Finally, the relatively high activity concentration of ^{137}Cs and the very small ratio $^{134}\text{Cs}/^{137}\text{Cs}$ values in all soil, grass and milk samples, far below the value of 1 (that was observed in atmospheric data), in combination with the absence of ^{134}Cs in most of these samples, is an indication that ^{137}Cs is the one remaining from Chernobyl and past global fallout rather than due to Fukushima fallout.

4. Dose Assessment

After the Fukushima accident a number of dose assessments have been carried out for the populations living in the north-west fallout zone of the Fukushima nuclear accident, by MEXT in Japan, DOE/NNSA in USA, IRSN in France [33, 34], with quite similar projected dose values. In the more affected regions the estimated projected doses reach particularly significant values, some of them even above 200 mSv, which are no longer in the range of “low doses” according to UNSCEAR 2000 definition [35]. The level of external projected doses in upcoming years is up to 4 Sv lifetime in the high-contaminated areas of 30 MBq m^{-2} of $^{137,134}\text{Cs}$.

On the contrary, the radioactive plume that reached European countries has only small amounts of radioactive isotopes. However, these isotopes, that were observed at low-level in the air boundary layer, were deposited by wet and dry deposition and have contaminated the land, and as a consequence the whole food chain. So the radioisotopes of cesium and iodine were found above their detection limits in all environmental samples but very far below levels of concern. Based on the observed values in our regions the total doses have been calculated.

The evaluation of the effective dose is done by the relation:

$$E = E_{est} + \sum_j h(g)_{j,ing} J_{j,ing} + \sum_j h(g)_{j,inh} J_{j,inh} < 1 \text{ mSv a}^{-1} \quad (1)$$

where,

E_{est} is the effective dose for exposure;

$J_{j,ing}$ and $J_{j,inh}$ are the intake activity (Bq) by ingestion and by inhalation of radionuclide j respectively;

$h(g)_{j,ing}$, $h(g)_{j,inh}$ (Sv Bq^{-1}) are the coefficients of committed dose for unit of intake by ingestion and/or by inhalation for the population of age group g , due to radionuclide j .

The evaluation of the committed dose is done using the maximum detected concentration values measured in the environment for ^{131}I , ^{134}Cs and ^{137}Cs in a hypothetical case when the radionuclide intakes are assumed to be constant for one year and taking into account the $h(g)$ coefficients for population of age younger than 1 year old and older than 17, as reported in the Italian Radioprotection Law [36]. In the calculation we assumed the annual individual usage factors for external exposure, inhalation and consumption of foods reported in NCRP-123 [37]. Even in this worst scenario, the values obtained with an extremely high conservative model confirm that the dose contributions of Fukushima fallout due to different pathways to Italian and Greek population were at least one order of magnitude less than the limit of 1 mSv a^{-1} . Furthermore and as a comparison, the derived dose equivalent from inhalation of airborne materials is very small compared to 1.25 mSv a^{-1} derived from inhalation of natural radiation sources, stemming mainly from radon 1.15 mSv a^{-1} and thoron 0.10 mSv a^{-1} [35].

5. Conclusions

Determination of ^{131}I , ^{137}Cs and ^{134}Cs in environmental samples were carried out in Thessaloniki (40°N) and in Milano (45°N) after the Fukushima accident during the period of March-May, 2011. All samples had very low levels of radioactive material, consistent with releases from the damaged Japanese nuclear reactors.

Activity concentrations of radionuclides in air measured at two different sites were almost similar taking into account the experimental errors. In these samples the $^{134}\text{Cs}/^{137}\text{Cs}$ ratio, throughout the measurement period, remained almost stable at about 1, value that is the fingerprint that these radionuclides are related to Fukushima disaster. On the contrary the relative high activity concentration of ^{137}Cs , the very small ratio $^{134}\text{Cs}/^{137}\text{Cs}$ far below the value of 1 in all soil, grass and milk samples, in combination with the absence of ^{134}Cs in most of these samples is an indication that the ^{137}Cs detected in these samples is related to the one remaining from Chernobyl and past global fallout in soil and as a consequence in grass and milk rather than due to Fukushima fallout.

The estimated committed doses for population related to the contributions of Fukushima fallout due to different pathways were at least one order of magnitude less of the limit of 1 mSv a^{-1} , even if the calculations are made using high conservative assumptions. In conclusion, radioisotopes of caesium and iodine were found above their detection limits in all environmental samples, but well below levels of concern.

References

1. IAEA, 02 June 2011. Fukushima Nuclear Accident Update Log. www.iaea.org/newscenter/news/tsunamiupdate01.html/.
2. Bolsunovsky A, Dementyev D (2011) Evidence of the radioactive fallout in the center of Asia (Russia) following the Fukushima Nuclear Accident. *J. of Environ. Radioactivity* 102: 1062-1064.
3. Bowyer TW, Biegalski SR, Cooper M, Eslinger PW, Haas D, Hayes JC, Miley HS, Strom DJ, Woods V (2011) Elevated radionuclides detected remotely following the Fukushima nuclear accident. *J. of Environ. Radioactivity* 102(7): 681-687.
4. Clemenza M, Fiorini E, Previtali E, Sala E (2012) Measurement of airborne ^{131}I , ^{134}Cs and ^{137}Cs due to the Fukushima reactor incident in Milan (Italy). *J. of Environ. Radioactivity* 114: 113-118.
5. Diaz LJ, Jaffe DA, Kaspar J, Knecht A, Miller ML, Robertson RGH, Schubert AG (2011) Arrival time and magnitude of airborne fission products from the Fukushima, Japan, reactor incident as measured in Seattle, WA, USA. *J. of Environ. Radioactivity* 102(11): 1032-1038.
6. Lozano RL, Hernández-Ceballos MA, Adame JA, Casas-Ruiz M, Sorribas M, San Miguel EG, Bolívar JP (2011) Radioactive impact of Fukushima accident on the Iberian Peninsula: Evolution and plume previous pathway. *Environ International* 37: 1259-1264.
7. Lujanienė G, Byčenkienė S, Ščiglo T, Gera M, Bartok J, Gažák M (2011) Radionuclides from the Fukushima Accident in Europe – Modelling the Air Mass Transport, Proceedings FSKD 2011, vol. 4, 26-28 July, 2011, Shanghai, China, 2775-2777.
8. Manolopoulou M, Vagena E, Stoulos S, Ioannidou A, Papastefanou C (2011) Radioiodine and radiocaesium in Thessaloniki, Northern Greece, due to the Fukushima nuclear accident. *J. of Environ. Radioactivity* 102(8): 796-797.
9. Masson O et al. (2011) Tracking of Airborne Radionuclides from the damaged Fukushima Dai-ichi nuclear reactors by European networks. *Environ. Sci. Technology* 45: 7670-7677.
10. Paatero J, Vira J, Siitari-Kauppi M, Hatakka J, Holmen K, Viisanen Y (2012) Airborne fission products in the high Arctic after the Fukushima nuclear accident, *J. of Environ. Radioactivity* 114: 41-47.
11. Pittauerová D, Hettwig B, Fischer HW (2011) Fukushima fallout in Northwest German environmental media. *J. of Environ. Radioactivity* 102(9): 877-880.

12. Povinec PP, Sýkora I, Holý K, Gera M, Kováčik A, Brest'áková L (2012) Aerosol radioactivity record in Bratislava/Slovakia following the Fukushima accident – A comparison with global fallout and the Chernobyl accident. *J. of Environ. Radioactivity* 114: 81-88.
13. Tositti L, Brattich E, Cinelli G, Previti A, Mostacci D (2012) Comparison of radioactivity data measured in PM₁₀ aerosol samples at two elevated stations in northern Italy during the Fukushima event. *J. of Environ. Radioactivity* 114: 105-112.
14. Arctic Monitoring and Assessment Programme (AMAP 2010), AMAP Assessment 2009: Radioactivity in the Arctic, Oslo, Norway, pp 2829, ISBN 13 9788279710592
15. Masson O, Piga D, Gurriaran R, D'Amico D (2010) Impact of an exceptional Saharan dust outbreak in France: PM10 and artificial radionuclides concentrations in air and in dust deposit. *Atmos. Environ.* 44: 2478–2486.
16. STAPLEX, www.staplex.com.
17. Ioannidou A, Papastefanou C (2006) Precipitation scavenging of ⁷Be and ¹³⁷Cs radionuclides in air. *J. Envir. Radioactivity* 85: 121-136.
18. Ioannidou A, Manenti S, Luigi G, Groppi F (2012) Fukushima fallout at Milano, Italy. *J. of Environ. Radioactivity* 114: 119-125.
19. Manolopoulou M, Stoulos S, Ioannidou A, Vagena E, Papastefanou C (2012) Radiation measurements and radioecological aspects of fallout from the Fukushima nuclear accident. *Journal of Radioanal. and Nucl. Chem.* 292(1): 155-159.
20. Firestone RB, Baglin CM, Chu SYF (1999) Table of Isotopes: 1999 Update on CD-ROM, 1999 Update, 8th Edition. Wiley-Interscience.
21. Gilmore GR (2008) Practical Gamma-ray Spectrometry, 2nd Edition, John Wiley & Sons, Ltd.
22. KEK (2011) KEK High Energy accelerator Research Organization, Measurement result of airborne nuclide and air radiation level in Tsukumba area (Japan): 3rd Report. 28 March, 2011. KEK.JP.
23. Kirchner G, Bossew P, De Cort M (2012) Radioactivity from Fukushima Dai-ichi in air over Europe; part 2: what can it tell us about the accident? *J. of Environ. Radioactivity* 114: 35-40.
24. IAEA (2006) Environmental consequences of the Chernobyl accident and their remediation: twenty years of experience”, RARS, VIENNA.
25. Paatero J, Hämeri K, Jaakkola T, Jantunen M, Koivukoski J, Saxén R (2010) Airborne and deposited radioactivity from the Chernobyl accident – a review of Investigations in Finland. *Boreal Environ. Res.* 15: 19-33.
26. Povinec P, Chudý M, Sýkora I, Szarka J, Pikna M, Holý K (1988) Aerosol radioactivity monitoring in Bratislava following the Chernobyl accident. *J. of Radioanal. and Nucl. Chem. Lett.* 126: 467-478.
27. Kritidis P, Florou H, Elrftheriadis K, Evangelou N, Gini M, Sotiropoulou M, Diapouli E, Vratolis S (2012) Radiative pollution in Athens, Greece due to the Fukushima nuclear accident. *J. of Environ. Radioactivity* 114: 100-104.
28. Lujaniene G, Byčenkienė S, Povinec PP, Gera M (2012) Radionuclides from the Fukushima Accident in the air over Lithuania: measurement and modeling approaches, *J. of Environ. Radioactivity* 114: 71-80.
29. GAEC (2011) Greek Atomic Energy Commission, Aghia Parakevi, 2011, Athens GR-15310. Bulletin 13 April 2011. www.eeae.gr.
30. Potiriadis C, Kolovou M, Clouvas A, Xanthos S (2011) Environmental radioactivity measurements in Greece following the Fukushima Daichi nuclear accident. *Rad. Prot. Dosimetry* 150: 441-447, doi:10.1093/rpd/ncr423.
31. Povinec PP, Sýkora I, Gera M, Holý K, Brest'áková L, Kováčik A (2013) Fukushima-derived radionuclides in ground-level air of Central Europe: a comparison with simulated forward and backward trajectories. *J. of Radioanal. and Nucl. Chem* 295: 1171-1176.
32. ATSDR (2011) Agency for Toxic Substances and Disease Registry http://www.atsdr.cdc.gov/csem/iodine/standards_regulations.html

33. MEXT, Ministry of Education, Culture, Sports, Science and Technologie, Japan.
<http://www.mext.go.jp/english/>
34. IRSN (2011) IRSN, Institut de Radioprotection et de Surete Nucleaire, Assessment on the 66th day of projected external doses from the nuclear accident in Fukushima, Report DRPH/2011-010, 27.05.2011
35. UNSCEAR 2000 United Nations Scientific Committee on the Effects of Atomic Radiation. Sources and effects of ionizing radiation, New York, United Nations.
36. Italian Government Legislative Decree n. 230 March 17 1995, as modified by D. Lgs. May 26 2000, n. 241 and May 9, 2001, n. 257 – Implementation of Directives 89/618/Euratom, 90/641/Euratom 92/3/Euratom and 96/29/Euratom on ionizing radiation.
37. NCRP 1996 “Screening models for releases of radionuclides to atmosphere, surface water and ground”, Recommendations of the National Council on Radiation Protection and Measurements, J REPORT No. 123 an 22, 1996.

Figure Captions

Fig. 1 Time dependence of the ^{131}I activity concentration measured in airborne particulates at the two laboratories in Milano and Thessaloniki

Fig. 2 Time dependence of the ^{137}Cs activity concentration measured in airborne particulates by the two laboratories in Milano and Thessaloniki

Fig. 3 Time dependence of the ^{134}Cs activity concentration measured in airborne particulates by the two laboratories in Milano and Thessaloniki

Fig. 1

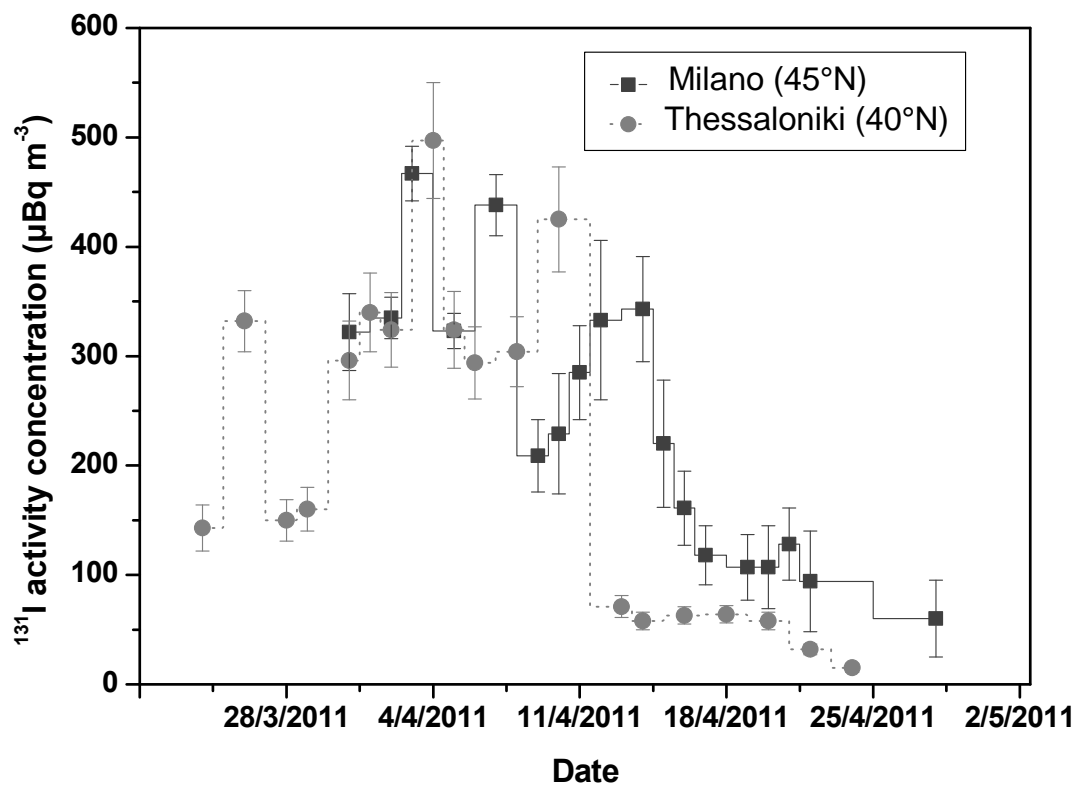


Fig. 2

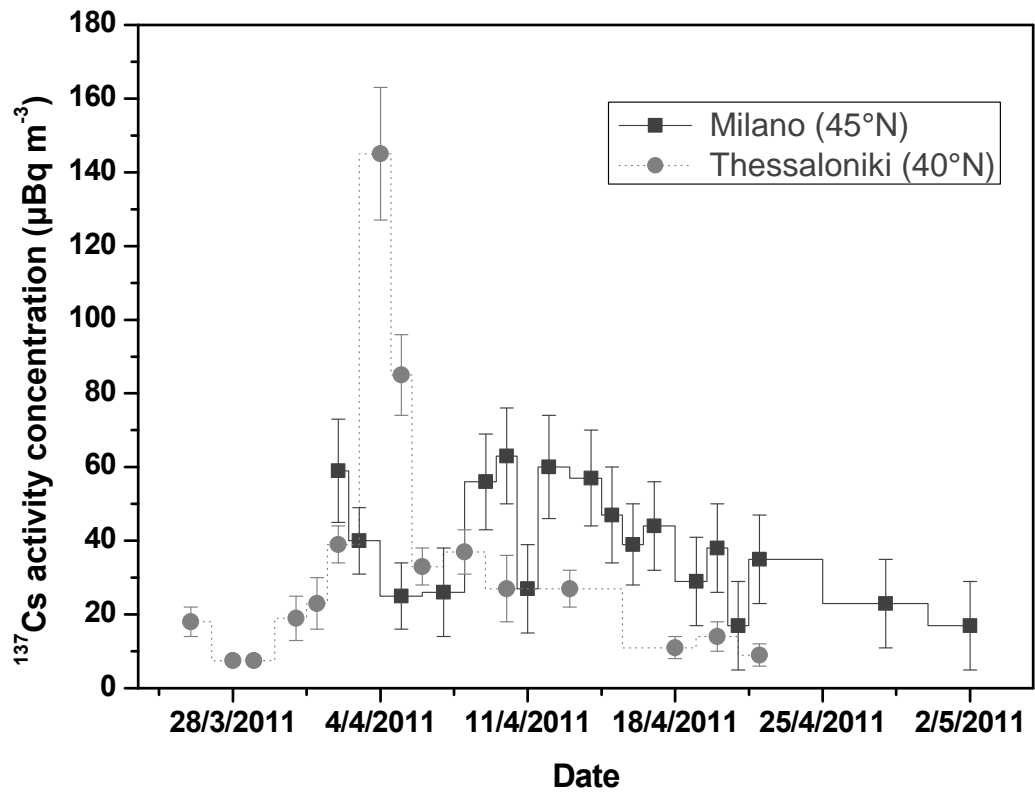


Fig. 3

